BIOENERGY/BIOFUELS/BIOCHEMICALS

# **Degradation of phenolic compounds by the lignocellulose deconstructing thermoacidophilic bacterium** *Alicyclobacillus Acidocaldarius*

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**Abstract** *Alicyclobacillus acidocaldarius*, a thermoacidophilic bacterium, has a repertoire of thermo- and acidstable enzymes that deconstruct lignocellulosic compounds. The work presented here describes the ability of *A. acidocaldarius* to reduce the concentration of the phenolic compounds: phenol, ferulic acid, ρ-coumaric acid and sinapinic acid during growth conditions. The extent and rate of the removal of these compounds were significantly increased by the presence of micro-molar copper concentrations, suggesting activity by copper oxidases that have been identified in the genome of *A. acidocaldarius*. Substrate removal kinetics was first order for phenol, ferulic acid, ρ-coumaric acid and sinapinic acid in the presence of 50  $\mu$ M copper sulfate. In addition, laccase enzyme assays of cellular protein fractions suggested significant activity on a lignin analog between the temperatures of 45 and 90 °C. This work shows the potential for *A. acidocaldarius* to degrade phenolic compounds, demonstrating potential relevance to biofuel production and other industrial processes.

**Keywords** *Alicyclobacillus acidocaldarius* · Thermophiles · Phenolics · Bioremoval

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## **Introduction**

*Alicyclobacillus acidocaldarius* is a thermoacidophilic bacterium capable of deconstructing lignocelluloses, hemicelluloses and celluloses as documented in research [[2,](#page-9-0) [17,](#page-10-0) [41,](#page-10-1) [42](#page-10-2)], reviews on the topic [[4,](#page-9-1) [5,](#page-9-2) [15\]](#page-9-3) and in components of several patent [\[33](#page-10-3)[–35](#page-10-4)]. Based on the ability of *A. acidocaldarius* to deconstruct hemicellulose and cellulose, we were interested in investigating its ability to deconstruct or lower the concentration of lignin derivatives, specifically the phenolic compounds: phenol, ferulic acid, ρ-coumaric acid and sinapinic acid, as these compounds represent the base derivatives of lignin groups.

Phenol and its derivatives are aromatic compounds that occur naturally in the environment, most commonly within lignocellulosic polymers, which are estimated to cycle 200 billion tons of carbon annually [[9\]](#page-9-4). The lignin polymer is the most significant component of lignocellulose by mass, and second only to cellulose in its contribution to living terrestrial biomass [[11](#page-9-5)]. Lignin is an amorphous, polyphenolic material constructed from the enzyme-mediated polymerization of three phenylpropanoid monomers, coniferyl, ρ-coumaryl and sinapyl alco-hols [[12\]](#page-9-6). Lignin also serves to protect the more simple polymers, hemicellulose and cellulose, from being quickly degraded by bacteria and fungi. Due to the amount of fixed carbon bound in lignin matrices globally, it has received a significant attention as a compound important to biofuel applications [[25](#page-10-5)]. To make biomass-to-fuel processes more economically feasible, it is important that technologies be developed that effectively remove lignin and its inhibitory derivatives, primarily phenolic compounds, from the energy rich, fermentable sugars that comprise the hemicellulose and cellulose.



In addition to being naturally ubiquitous, phenol and phenolic compounds are also produced during industrial processes, including petroleum refining, plastic manufacturing, resin production, pesticide production, steel manufacturing and the production of paints and varnish [\[3](#page-9-7), [19\]](#page-10-6). Because phenolic compounds are water soluble, they are highly mobile in waste waters generated during these processes [\[10](#page-9-8)], and may be distributed to downstream river, lake, and soil sites, where they can accumulate to higher concentrations and cause ecological concerns [[13](#page-9-9)]. Because of this, several techniques have been developed to remove phenol and its derivatives from industrial effluents and contaminated waters with bioremediation receiving the most attention due to its low environmental impact and relatively low cost [[26](#page-10-7)]. To this end, the removal of phenol has been observed using bacterial and fungal species including *Candida tropicalis* [\[30](#page-10-8)], *Acinetobacter calcoaceticus* [\[24\]](#page-10-9), *Alcaligenes eutrophus* [[18](#page-10-10)], *Pseudomonas putida* [\[22\]](#page-10-11) and *Burkholderia cepacia* G4 [[31](#page-10-12)].

With respect to the degradation and removal of phenol and phenolic compounds, genes for five putative multi-copper oxidases (GenBank Accession Numbers: ACV 59461.1; ACV 59462.1; ACV 58146.1; ACV 58141.1 and ACV 59388.1) have been identified in the sequenced genome of *A. acidocaldarius* [[21\]](#page-10-13). These genes and select properties are shown in the supplemental files. Multi-copper oxidase such as laccase and laccase-like multi-copper oxidases (LMCO) are copper-containing genes that have been shown to act as polyphenol or phenol oxidases [\[27](#page-10-14)]. The presence of the genes for these enzymes prompted the addition of copper to certain media sets to determine any effects on the removal of phenolic compounds by *A. acidocaldarius*.

The work presented here builds on these and other efforts by reporting the ability of *A. acidocaldarius*, a thermoacidophile, to degrade phenol and its common derivatives, ferulic acid, ρ-coumaric acid and sinapinic acid under aerobic growth conditions. The findings presented suggest that *A. acidocaldarius* and/or enzymes produced by *A. acidocaldarius* may have applications in removing phenolic compounds from processes or waste streams at high temperatures and low pHs, conditions common for many bioenergy and industrial processes, including biomass preconversions, lignin deconstruction/solubilization and the mediation of potentially inhibitory phenolic compounds that are produced during the conversion of lignocellulose to fermentable sugars during biofuel applications.

# **Materials and methods**

# **Culture medium and inoculum preparation**

Chemicals were obtained from Fisher Chemical (Fairlawn, NJ, USA). *A*. *acidocaldarius* (ATCC 27009) was cultured in a basal mineral medium that contained the following (in  $g L^{-1}$ ): CaCl<sub>2</sub>·2H<sub>2</sub>O, 0.25; MgSO<sub>4</sub>·7H<sub>2</sub>O, 0.5; (NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub>, 1.3; Fe(III)  $\cdot$ EDTA, 0.047; KH<sub>2</sub>PO<sub>4</sub>, 3.0 and xylose, 4.17 (30 mM). The xylose was withheld from substrate-free controls. In addition, 1 mL  $L^{-1}$  each of trace mineral stock solutions A, B, and C was added. Stock solution A contained (in g L<sup>-1</sup>): MgCl<sub>2</sub>, 25.0; CaCl<sub>2</sub>·2H<sub>2</sub>O, 6.6; H<sub>3</sub>BO<sub>3</sub>, 0.58; Fe(III)Cl<sub>3</sub>·6H<sub>2</sub>O, 5.0; CoNO<sub>3</sub>·6H<sub>2</sub>O, 0.05; and NiCl<sub>2</sub>·6H<sub>2</sub>O, 0.02. Stock solution B contained (in g  $L^{-1}$ ):  $MnSO_4·H_2O$ , 2.0;  $ZnSO_4·7H_2O$ , 0.5; and  $Na_2MoO_4$ , 0.025. Stock solution C was a vitamin solution that contained (in g L<sup>-1</sup>): pyridoxine hydrochloride, 0.08; folic acid, 0.012; thiamine hydrochloride, 0.13; riboflavin, 0.042; nicotinamide, 0.084; ρ-aminobenzoate, 0.088; biotin, 0.01; cyanocobalamin, 0.0004; p-pantothenic acid calcium, 0.086; myo-inositol, 0.021; choline bromide, 0.053; orotic acid sodium salt, 0.021; and spermidine, 0.1. After the basal salts medium and stock solutions were mixed, the medium pH was adjusted to 4.0 using 6 N HCl and the medium was filtered  $(0.2 \mu m)$  under sterile conditions. Phenolic compounds and copper sulfate (as a potential enzyme co-factor) were added prior to filtration. The medium was then incubated at 60 °C for up to 1 h to solublilize the phenolic compounds.

Prior to each experiment, *A*. *acidocaldarius* cells preserved at −80 °C in 1-mL vials at 10 % dimethyl sulfoxide (DMSO) were introduced as the initial inoculum. Cultures were grown in 125-mL Erlenmeyer flasks (40 mL medium volume), fitted with foam stoppers, and shaken at 120 rotations per minute (rpm) in an Innova 4230 temperature controlled incubator (New Brunswick, Edison, NJ, USA) at 60 °C. These initial cultures were allowed to grow into the late-exponential phase, when 0.4 mL aliquots were collected, concentrated via centrifugation (7000×*g*), and washed in substrate-free medium, before being inoculated into fresh medium. This growth and re-inoculation process was repeated two times to allow the cells to adapt to the experimental conditions. Cell growth was monitored by measuring optical density at 595 nm using an UV–Vis V160u spectrophotometer (Shimadzu, Columbia, MD, USA). All experiments were performed in triplicate and average values and 95 % confidence intervals were calculated.

## **Measurement of phenolic compounds**

An Agilent 7100 Capillary Electrophoresis (Foster City, CA, USA) instrument equipped with a diode-array detector was used to measure the concentrations of phenol, ferulic acid, ρ-coumaric acid and sinapinic acid. Data acquisition and analysis were accomplished using Agilent ChemStation software. Methods developed by Bonoli et al. [[6\]](#page-9-10) were used, with the exception that the conditioning steps using HPCE-grade (Fisher Optima) and buffer solution were done for 5 min, rather than 2 min. This change resulted in more defined peaks when several consecutive samples were run. The running buffer used was 45 mM sodium tetraborate (pH 9.6). Samples were injected at the anodic end in a low pressure mode (3.4 kPa) for 3 s. Separation was carried out using a voltage of 27 kV for 10 min with a capillary temperature of 30 °C. The resulting current was 110 μA. Before each injection, the capillary was rinsed in high pressure mode (137.9 kPa) with 0.1 N HCl for 2 min, HPCE-grade water for 2 min, and then re-equilibrated with the running buffer for 2 min. After each cycle, the capillary was rinsed with HPCE-grade water for 2 min. All rinses were conducted at 30 °C. Phenol, ferulic acid, ρ-coumaric acid and sinapinic acid eluted at approximately 7.0, 5.2, 5.9 and 3.8 min, respectively. Sample concentrations were determined from correlations calculated using analytical grade standards. Using these methods, the detection limits for phenol, ferulic acid, ρ-coumaric acid and sinapinic acid were approximately  $10 \mu M$ .

#### **Collection of protein fractions**

To isolate protein fractions, cells grown under conditions identical to those used in the phenolic removal experiments were harvested by centrifuging 40 mL-culture aliquots at  $7000 \times g$  for 15 min. The protein isolation method described here was adopted from Zambrano et al. [\[40](#page-10-15)]. The supernatant was decanted and the pellet was washed with 5 mL of a 25 mM phosphate buffered saline (PBS) solution. This process was repeated twice before 5 μL of DNase/RNase and 20 μL of a protease inhibitor cocktail were added to the samples, which were then frozen using liquid nitrogen, and allowed to thaw. The freeze–thaw application was repeated twice. Samples were then transferred to a 15 mL conical-bottom tube and sonicated twice for 1 min. Visual inspection using a light microscope at  $1000 \times$  suggested significant cell lysis. The samples were allowed to incubate at room temperature for 15 min, and then centrifuged at  $15,000 \times g$  for 35 min at 4 °C. The supernatant was collected and, if necessary, saved at −80 °C. The pellet was then re-suspended in a 0.5 M NaCl solution with sodium acetate added to a concentration of 20 mM. DNase/RNase and protein inhibitor cocktails were added at 5 and 20  $\mu$ L, respectively. This solution was centrifuged at 15,000×*g* for 25 min and the supernatant was added to the fraction collected previously. Collectively, this fraction was considered to predominantly contain soluble proteins, such as those often concentrated in the cytoplasm. The pellet was again re-suspended in 200  $\mu$ L of 0.4 % triton X-100 in 25 mM PBS and centrifuged for 25 min at 15,000×*g*. The resulting supernatant was collected and considered to predominantly contain membrane-associated proteins. The remaining

pellet was re-suspended in 200 μL of PBS and collected as the insoluble, membrane-bound, fraction. The protein concentrations in these fractions were measured using a Bradford assay kit (Bio-Rad, Hercules, CA, USA), and diluted to approximately 10  $\mu$ g  $\mu$ L<sup>-1</sup> in the reaction buffers. It should be noted that this method was used to affect some degree of fractionation between cytoplasmic/soluble proteins and membrane-bound proteins. The fractions are denoted as such, but this was a qualitative effort and the original cellular location of the separated proteins was not confirmed.

#### **Enzyme assays**

Each of the enzyme reactions was conducted at 60 °C and stirred at 200 rpm in a heated, temperature controlled cell within an Evolution 600 UV–vis spectrophotometer (Thermo Scientific, Waltham, MA, USA). The absorbance of the various assay-dependent oxidation products was measured with time, and data were collected and analyzed using VisionPro software (Visionpro, Calgary, Canada). The following assays were used:

#### *Laccase activity*

The enzyme solution (0.1 mL) was added to 0.1 mL sodium acetate (0.1 M, pH 5.0) and 0.8 mL of 0.3  $\%$  (w/v) 2,2′-azino-bis (3-ethylbenzothiazoline-6-sulphonic acid) (ABTS). The oxidation product of ABTS was monitored with time at 420 nm.

#### *Manganese peroxidase*

0.5 mL of the enzyme solution was added to 0.1 mL of 0.1 % (w/v) phenol red, 0.1 mL of 0.25 M sodium lactate, 0.2 mL of 0.5  $\%$  (w/v) bovine serum albumin, 0.05 mL of 2 mM  $MnSO<sub>4</sub>$  and 0.05 mL of 100 mM  $H<sub>2</sub>O<sub>2</sub>$  in 20 mM sodium succinate (pH 4.5). The oxidation of phenol red was monitored at 610 nm.

#### *Lignin peroxidase*

0.5 mL of the enzyme solution was added to 0.5 mL  $H_2O_2$ , 0.5 mL of 10 mM veratryl alcohol and 1.0 mL of 125 mM sodium tartrate (pH 3.0). The reaction mixture was monitored at 310 nm to measure  $H_2O_2$  oxidation.

#### **Modeling copper solubility and speciation**

Visual MINTEQ (version 2.53) was used to predict the speciation of copper sulfate that was added to sets of growth media as a potential enzyme co-factor. This modeling was done to approximate the amount of soluble and free-form copper that might be biologically available to the cells during growth. To determine the likely speciation of the media components, MINTEQ solved the thermodynamic equilibrium solutions using activities from the Debye–Huckel equation and the default MINTEQA2 thermodynamic database. The temperature was set to 60 °C and the free-proton concentration was calculated from the pH, which was fixed at 4.0. This method resulted in a prediction that over 99 % of the copper added would exist in the free-divalent form with respect to abiotic considerations.

# **Statistics and calculations**

For each experiment, tests were performed in triplicate and average values and 95 % confidence intervals were calculated. The 95 % confidence intervals were calculated using a t-distribution assumption, which resulted in somewhat larger error bars than what would have been observed had a normal distribution of data variance been assumed. This is a more conservative estimate, and likely more accurate assumption, based on the limited number of replicates.

# **Results**

# **Effect of copper concentration on cell growth and phenolic compound removal**

When *A. acidocaldarius* was grown with xylose as the carbon and energy source, copper in the growth medium appeared to extend the lag phase of growth from between three and six hours without copper, to over nine hours when 100 μM copper was added (Fig. [1\)](#page-3-0). In addition, the rate of cell growth and total cell yield decreased by approximately 20–30 % (calculations not shown). However, when copper

<span id="page-3-0"></span>

was added to a concentration of 10  $\mu$ M, there appeared to be no adverse effect on the total cell yield or rate of growth. Adding copper to 25 or 50  $\mu$ M decreased total cell yields by less than 10 %, but did not have an appreciable effect on the rate of cell growth. The time required for the cultures to reach stationary phase ranged from 24 h at copper concentrations of 0 and 10  $\mu$ M to approximately 30 h at copper concentrations of 25, 50 and 100  $\mu$ M.

The amount of copper in the culture media affected the rates and extent of phenolic compound removal by *A. acidocaldarius* (Fig. [2](#page-4-0)). In the absence of copper, phenol was degraded from a concentration of  $0.52 \pm 0.01$  mM to  $0.18 \pm 0.03$  mM in 24 h (Fig. [2a](#page-4-0)). However, when copper was added to an initial concentration of 100 μM, phenol was reduced to a concentration of  $0.10 \pm 0.02$  mM in 6 h, and was completely removed (to below the detection limit) within 24 h. Copper had similar effects on the removal of ferulic, ρ-coumaric and sinapinic acids. For example, no statistically significant removal was measured in cultures containing ferulic acid when copper was absent; however, when copper was added to a concentration of  $100 \mu M$ , ferulic acid concentrations decreased from  $0.50 \pm 0.03$  mM to  $0.29 \pm 0.01$  mM over 48 h (Fig. [2b](#page-4-0)). Similar observations were made of cultures containing ρ-coumaric and sinapinic acid (Fig. [2c](#page-4-0), d), although some removal of sinapinic acid was observed in the absence of copper, from  $0.51 \pm 0.01$  mM to  $0.43 \pm 0.01$  mM over 48 h. Experiments using substrate-free (live cells) and heat-killed cells with and without copper showed no statistically significant decrease in phenolic compound concentrations (data not shown), suggesting that the decreases in phenolic concentrations are not due to cellular sorption.

Initial removal rates were calculated from mass balance equations and plotted against the copper concentrations used in each experiment (Fig. [3\)](#page-5-0). The relationships







<span id="page-4-0"></span>**Fig. 2** Removal of **a** phenol, **b** ferulic acid, **c** ρ-coumaric acid and **d** sinapinic acid by a growing culture of *Alicyclobacillus acidocaldarius* in the presence of copper  $(0, 10, 25, 50$  or  $100 \mu M$ ). The phenolic

compounds were each added at an initial concentration of 0.5 mM. The values shown represent the averages of triplicate experiments ±95 % confidence intervals

resembled first-order Michaelis–Menten substrate kinetics, with the exception that relatively low reaction rates were observed in the absence of copper. The  $V_{\text{max}}$  and  $K_{\text{m}}$ values for each phenolic compound (phenol, ferulic acid, ρ-coumaric acid and sinapinic) were determined graphically with respect to medium copper concentration using Lineweaver–Burk plots, and are shown in Table [1.](#page-5-1) While  $K<sub>m</sub>$  and  $V<sub>max</sub>$  were significantly greater for phenol than the phenolic acids, they were statistically equivalent among ferulic, ρ-coumaric and sinapinic acid. Of note, these data show that the removal rates of phenolic compounds increase with copper concentration up to approximately  $50 \mu$ M. Using higher copper concentrations did not result in statistically higher rates of removal.

# **Effect of initial concentrations of phenolic compounds on cell growth and on the removal of phenolic compounds**

The extent and Michaelis–Menten relationships of phenolic compound removal were observed for *A.acidocaldarius* with respect to the initial concentrations of phenolic compounds (Figs. [4](#page-6-0), [5\)](#page-7-0). In addition, the inhibitory effects on the growth of *A. acidocaldarius* cultured in the presence of varying concentrations of phenol, ferulic acid, ρ-coumaric acid and sinapinic acid (Fig. [4](#page-6-0)) are described here. As with copper, the data show that phenol increases the time of lag-phase growth, from less than 6 h in the absence of phenol to between 6 and 12 h when phenol is present in concentrations up to 5 mM. In addition, there is a statistically significant effect on the total cell yield and growth rate when phenol is added to a concentration of 5 mM, but little or no effect when added to lower concentrations. Ferulic acid, ρ-coumaric acid and sinapinic acid also increase the duration of lag-phase growth from under 6 h in the absence of these compounds to over 18 h when these compounds are added to a concentration of  $750 \mu$ M. Although there is little effect of these compounds on the total cell yield, the rate of growth decreased steadily with increasing initial concentrations.

The rate and extent of removal for each of the compounds tested increased with initial concentration. In each of these experiments, the copper concentration was set at 50  $\mu$ M and, like copper, the relationship between phenolic compound removal and initial concentration appears to be first order in nature. When added to concentrations of 0.5, 1.0 or 2.0 mM, phenol was completely degraded within 24 h. When added to an initial concentration of 5.0 mM, it was degraded to  $0.95 \pm 0.20$  mM after 24 h. Between 25 and 50 % of the



<span id="page-5-0"></span>**Fig. 3** Removal of **a** phenol and **b** ferulic acid, ρ-coumaric acid and sinapinic acid by *Alicyclobacillus acidocaldarius* in the presence of copper. The initial concentration for each of the phenolic compounds was 0.5 mM. The values shown represent the averages of triplicate experiments ±95 % confidence intervals

<span id="page-5-1"></span>**Table 1**  $K_m$  (half-saturation constant) and  $V_{\text{max}}$  (maximum reaction rate) as determined by Michaelis–Menten kinetics in a Linewever Burk plot, with respect to the initial copper concentration

	$K_{\rm m}$ (mM)	$V_{\text{max}}$ (mmol $h^{-1}$ )
Phenol	$11.2 \pm 2.8$	$0.145 \pm 0.306$
Ferulic acid	$6.03 \pm 2.04$	$0.006 \pm 0.002$
p-Coumaric acid	$3.60 \pm 1.87$	$0.006 \pm 0.002$
Sinapinic acid	$5.51 \pm 1.80$	$0.008 \pm 0.002$

These data describe the dependency of enzyme activity on the copper concentration (half-saturation constant) and the maximum potential reaction rate of the enzyme (maximum reaction rate) with increasing copper concentration. These parameters are one way to describe the importance of copper concentrations in the reactions that remove each of the compounds tested here (see Fig. [3](#page-5-0)). The initial concentration of each compound was 0.5 mM. The values shown represent the averages of triplicate experiments  $\pm$  95 % confidence intervals. The confidence intervals suggest that the  $K<sub>m</sub>$  and  $V<sub>max</sub>$  data for phenol are clearly significantly higher than those for the phenolic acids. However, there is no statistically significant difference in the  $K<sub>m</sub>$  or  $V<sub>max</sub>$ data among the phenolic acids

initial ferulic acid, ρ-coumaric acid and sinapinic acid were degraded over a 36 to 72 h period when added to initial concentrations between 250 and 750 μM. However, when added to an initial concentration of 100 μM, each of these compounds was nearly completely degraded within 36 h. Phenol, ferulic acid, ρ-coumaric acid and sinapinic acid show firstorder relationships with respect to their initial concentration (Fig. [5\)](#page-7-0). The  $K<sub>m</sub>$  and  $V<sub>max</sub>$  values for phenolic removal were calculated with respect to the phenolic concentrations, as described previously, and are shown in Table [2.](#page-7-1) Again, the  $K<sub>m</sub>$  and  $V<sub>max</sub>$  values for phenol removal were significantly higher than those for ferulic, ρ-coumaric and sinapinic acid, which were not statistically different from each other within 95 % confidence.

#### **Enzyme assay activity**

The enzyme activities of collected protein fractions were calculated from assays designed to measure substrate oxidation by laccase-like, lignin peroxidase-like and manganese peroxidase-like enzymes (Table [3](#page-8-0)). These fractions were collected using methods designed to isolate the cytoplasmic/soluble, membrane-associated, and membranebound proteins. However, as the proteins in these fractions were not characterized, their true origin should be considered unknown, but with likely affinities towards either cytoplasmic origins, membrane-associated origins or membrane-bound origins. In each case, the addition of copper to the laccase assay increased the enzyme activity several fold. This activity was largely relegated to the membraneassociated and membrane-bound fractions.

Because of the fundamental mechanisms attributed to lignin and manganese peroxidases, these assays were carried out in the absence of copper. Some activity was measured in the soluble fraction, and a very small, nearly statistically insignificant amount was observed from the membrane-associated fraction. No activity was observed in the membrane-bound fraction for either lignin or manganese peroxidase.

Laccase activity from the membrane-associated fraction was the most significant activity measured, and as this was a copper-dependent effect, it may be relevant to the removal of phenolics observed in the whole-cell experiments. Because of this, the laccase assays were repeated with the membrane-associated fraction between 45 and 90 °C to measure the effect of temperature on this reaction. Activities (normalized to the maximum observed activity) of  $0.32 \pm 0.15$ ,  $1.00 \pm 0.07$ ,  $0.97 \pm 0.04$ , and  $0.23 \pm 0.08$ at 45, 60, 75 and 90 °C, respectively were measured (data shown in the supplemental material).

# **Discussion**

*A. acidocaldarius* is able to reduce the concentration of phenolic compounds as a function of both initial copper



<span id="page-6-0"></span>**Fig. 4** *Alicyclobacillus acidocaldarius* growth and removal of phenol (**a**, **e**), ferulic acid (**b**, **f**), ρ-coumaric acid (**c**, **g**) and sinapinic acid (**d**, **h**) in the presence of 50 μM copper. The values shown represent the averages of triplicate experiments  $\pm$ 95 % confidence intervals

and initial phenolic compound concentrations. Relevant to these findings, and the potential applications, the data show that although some growth inhibition results from the applicable concentrations of these compounds, this inhibition does not appear to interfere significantly with the removal of phenol, ferulic acid, ρ-coumaric acid or sinapinic acid.

In addition, the majority of the substrate is metabolized before the cultures reached the late-exponential growth phase. These two findings suggest that the enzymes needed to carry out the removal of phenolic compounds were not limited by the relatively low cell concentrations and slowed growth during the initial conditions tested here. The



<span id="page-7-0"></span>**Fig. 5** Reaction rates for the removal of **a** phenol and **b** ferulic acid, ρ-coumaric acid and sinapinic acid by *Alicyclobacillus acidocaldarius* in the presence of copper at 50  $\mu$ M. The values shown represent the averages of triplicate experiments  $\pm 95$  % confidence intervals

<span id="page-7-1"></span>**Table 2**  $K_m$  (half-saturation constant) and  $V_{max}$  (maximum reaction rate) as determined by Michaelis–Menten kinetics in a Linewever Burk plot, with respect to initial substrate concentration

	$K_{\rm m}$ (mM)	$V_{\text{max}}$ (mmol $h^{-1}$ )
Phenol	$1.01 \pm 0.13$	$0.460 \pm 0.031$
Ferulic acid	$0.37 \pm 0.09$	$0.017 \pm 0.004$
p-Coumaric acid	$0.21 \pm 0.06$	$0.014 \pm 0.003$
Sinapinic acid	$0.20 \pm 0.05$	$0.013 \pm 0.003$

These data describe the dependency of enzyme activity on the substrate concentration (half-saturation constant) and the maximum potential reaction rate of the enzyme (maximum reaction rate) with increasing substrate concentrations. These parameters are one way to describe the effects of substrate concentration on the reactions that remove the compounds tested here (see Fig. [5](#page-7-0)). The initial copper concentration in each test was 50  $\mu$ M. The values shown represent the averages of triplicate experiments  $\pm$  95 % confidence intervals. The confidence intervals suggest that the  $K<sub>m</sub>$  and  $V<sub>max</sub>$  data for phenol are clearly significantly higher than those for the phenolic acids. The  $K<sub>m</sub>$  value for ferulic acid also appears to be higher than the  $K<sub>m</sub>$  values for ρ-Coumaric acid and Sinapinic acid. However, there is no statistically significant difference in the  $V_{\text{max}}$  data among the phenolic acids

first-order dependence on phenolic concentrations (Fig. [4\)](#page-6-0) supports the suggestion that, under the conditions tested here, the removal rate of phenolic compounds is not limited by enzyme concentrations, or factors affecting enzyme concentrations, such as cellular growth rates and total cell numbers.

The removal rates and kinetic parameters measured here are greater to, or comparable to, several previous studies of phenolic compound removal [\[7](#page-9-11)] and are also comparable to values observed during the oxidation of BTEX compounds [\[37](#page-10-16)]. Combined with the thermos- and acidic-tolerance of *A. acidocaldarius* and its enzymes, these results are relevant to several biotechnological applications, specifically those related to the feedstock preconversions and biochemical conversions of biomass feedstocks where lignin-derived phenolics may inhibit downstream microbial activities. Within this study, the greater kinetic parameters and higher removal rates observed for phenol, compared to ferulic acid, ρ-coumaric acid and sinapinic acid, are not surprising, as phenol offers less steric hindrance to direct enzyme oxidation. Likewise, the similarity in parameters and removal rates observed for ferulic acid, ρ-coumaric acid and sinapinic acid are not surprising, as these are structurally similar compounds. However, these comparisons are relevant nonetheless, as there is a lack of literature directly comparing the removal, and removal kinetics, of these compounds.

The observed first-order dependence on copper (Fig. [3\)](#page-5-0) is interesting, and may suggest that *A. acidocaldarius* does not accumulate copper to a significant extent; in the manner that other microorganisms have been observed to concentrate necessary co-factors, sometimes by several orders of magnitude [\[28](#page-10-17)]. Conversely, this observation suggests that at least some copper-dependent oxidases in *A. acidocaldarius* simply replace copper after it oxidizes the substrate, in the case that the enzyme transfers electrons directly to the substrate. Another possibility is that the genetic translation of the relevant enzymes has pseudo-first-order dependencies on copper concentrations. Such dependencies on protein translation have been reported previously for copper oxidases, although no kinetic data were reported [[39\]](#page-10-18).

In addition to the whole-cell experiments, enzyme assays were carried out using various protein fractions and substrate analogs. It is interesting to note that, in the cytoplasmic and membrane-associated fractions, some non-copper-dependent activity was observed for the lignin peroxidase assays, which measured the oxidation of the relatively simple 2,4-dichlorophenol compound. This activity could explain how phenol and, to a lesser extent, ferulic acid, ρ-coumaric acid and sinapinic acid, were degraded in the absence of copper. Conversely, the highest activities were observed in the membrane-associated and membrane-bound fractions in the laccase assays. These activities were highly copper-dependent, suggesting that most of the phenolic removal observed may have been due to LMCOs, such as those suggested by genes in the genome of *A. acidocaldarius* (shown as a table in supplemental

<span id="page-8-0"></span>**Table 3** Measured activities for the laccase, lignin peroxidase and manganese peroxidase assays  $\pm$  95 % confidence intervals using various protein fractions



The values reported represent units of activity, and are equivalent to 1  $\mu$ mol  $\mu$ g<sup>-1</sup> min<sup>-1</sup>. The values shown represent the averages of triplicate experiments  $\pm$  95 % confidence intervals. Note that the total activity w/copper for the lignin peroxidase and manganese peroxidase tests is listed as 'Not Available,' as each of the fractions tested w/copper yielded activities below the detection limit

*BDL* below detection limit, which was approximately 1 unit of activity

<span id="page-8-1"></span>



files). In silico analysis of the protein products from three of the five genes analyzed showed the presence of signal peptides indicating transfer of these proteins to the exterior of the cell. This analysis provides additional support for the higher aromatic monomer degradation activities for the membrane-bound and membrane-associated fractions.

Assimilation of phenol and lignin-associated aromatic monomers by *A. acidocaldarius* for growth is further supported by the presence of genes for the *meta*-fission pathway for phenol degradation (Table [4\)](#page-8-1). In addition, Aaci\_1623 encodes an enzyme involved in metabolism of aromatic compounds and may be involved in phenol metabolism. This metabolism proceeds through the 4-oxalocrotonate branch as opposed to the hydrolytic branch, a characteristic seen in other bacteria, such as *Comamonas testosteroni* strain TA441 [\[1](#page-9-12)]. In contrast to phenol hydrolase catalyzed oxidation of phenol to catechol in *C. testosteroni*, phenol oxidation by *A. acidocaldarius* is catalyzed by one or more of the putative multi-copper oxidases.

Using this proposed mechanism, phenol is indirectly hydroxylated by the multi-copper oxidase [[36\]](#page-10-19), followed by ring cleavage and further breakdown into compounds that feed into the tricarboxylic acid cycle by other enzymes of the *meta*-fission pathway for phenol metabolism.

Laccase and LMCOs have great biotechnological potential for converting lignin in plant biomass to less structurally complicated polyphenols or aromatic monomers [\[14](#page-9-13)], as well as applications in the forest products industry [\[38\]](#page-10-20). In addition, like the lignin and manganese peroxidases, laccases are also thought to oxidize phenolic lignin residues [[32\]](#page-10-21), making such assays applicable in the remediation of certain processes, such as those in the textile industry. In line with such potential applications, the research presented here suggests that putative LMCOs of *A. acidocaldarius* do in fact degrade phenol and the phenolic compounds that are derivatives of compounds integral to lignocellulosic structures.

Previously, laccase genes and proteins have been primarily studied in mesophilic bacteria, including species of *Bacillus*

[\[29\]](#page-10-22), *Pseudomonas* [\[16](#page-10-23)] and *Streptomyces* [\[23\]](#page-10-24). In addition, thermo- and hyperthermophilic laccases have been isolated from prokaryotes, including Eubacteria, such as *Thermus thermophilus*, *Thermobifida fusca* and *Thermobaculum terrenum* [\[8](#page-9-14)], and *Archaea* such as *Pyrobaculum aerophilum* [\[20\]](#page-10-25). The work presented here, notably the temperature range over which the studied enzyme fractions were active, builds on such results and implies a broad and industrially relevant temperature range for biotechnological applications for the enzymes catalyzing this activity. Specifically, the temperature range tested includes conditions that may be used in the preconversions of lignocellulosic feedstocks where organic extractives, including phenolics, may be liberated. Of note, previous efforts have identified glycosyl hydrolase enzymes from *A. acidocaldarius* that showed activity across a similar temperature range (unpublished data).

These findings contribute to an important research need to identify and characterize biomass degrading enzymes that are active under conditions often encountered in the biorefinery of lignocelluloses (i.e., high temperature, low pH). The use of these specialized enzymes in the processes may enable the mineralization of biomass to occur at higher temperatures (less cooling of substrate) and lower pH (less neutralization), saving money in both capital and operating costs. In addition to the possible deconstruction of biomass, the removal of extractable organics, such as phenolics, may benefit downstream conversion processes.

# **Conclusions**

The present study shows the potential for the thermoacidophile *A. acidocaldarius* to degrade phenolic compounds found in lignin demonstrating potential relevance to biofuel production and other industrial processes. The majority of the observed removal was dependent on added copper, both in whole-cell and enzyme assay experiments. This suggests that putative multi-copper oxidases identified in the genome of *A. acidocaldarius* demonstrated laccase-like activity and may play an important role in the degradation of these lignin byproducts and constituents. Enzyme assays demonstrated that the majority of the activity was due to fractions associated with membrane-associated or membrane-bound proteins. Finally, observed enzyme activities from 45 to 90 °C show that *A. acidocaldarius* may possess laccase-like enzymes applicable to biotechnological processes over a wide range of temperatures.

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#### **Compliance with ethical standards**

**Conflict of interest** The authors declare that they have no conflict of interest.

**Human participants and/or animals** This article does not contain any studies with human participants or animals performed by any of the authors.

**Informed consent** Informed consent was obtained from all individual participants included in the study.

# **References**

- <span id="page-9-12"></span>1. Arai H et al (2000) Arrangement and regulation of the genes for meta-pathway enzymes required for degradation of phenol in *Comamonas testosteroni* TA441. Microbiology 146:1707–1715
- <span id="page-9-0"></span>2. Bai Y, Wang J, Zhang Z, Yang P, Shi P, Luo H, Meng K, Huang H, Yao B (2010) A new xylanase from thermoacidophilic *Alicyclobacillus* sp. A4 with broad-range pH activity and pH stability. J Ind Microbiol Biotechnol 37:187–194
- <span id="page-9-7"></span>3. Banyopadhyay S, Soumya H (1998) Briefing paper on forest policy: community stewardship and management. Terracotta Summit
- <span id="page-9-1"></span>4. Bhalla A, Bansal N, Kumar S, Bischoff KM, Sani RK (2012) Improved lignocellulose conversion to biofuels with thermophilic bacteria and thermostable enzymes. Bioresour Technol 128:751–759
- <span id="page-9-2"></span>5. Blumer-Schuette SE, Brown SD, Sander KB, Bayer EA, Kataeva I, Zurawski JV, Conway JM, Adams MWW, Kelly RM (2014) Thermophilic lignocellulose deconstruction. FEMS Mirobiol Rev 38:393–448
- <span id="page-9-10"></span>6. Bonoli M, Montanucci M, Toschi TG, Lercker G (2003) Fast separation and determination of tyrosol, hydroxytyrosol and other phenolic compounds in extra-virgin olive oil by capillary zone electrophoresis with ultraviolet-diode array detection. J Chromatogr A 10:163–172
- <span id="page-9-11"></span>7. Boyd TJ, Carlucci AF (1993) Degradation rates of substituted phenols by natural populations of marine bacteria. Aquat Toxicol 25:71–82
- <span id="page-9-14"></span>8. Brander S, Mikkelsen JD, Kepp KP (2015) TtMCO: a highly thermostable laccase-like multicopper oxidase from the thermophilic *Thermobaculum terrenum*. J Mol Cat B Enzym 112:59–65
- <span id="page-9-4"></span>9. Chandel AK, Singh OV (2011) Weedy lignocellulosic feedstock and microbial metabolic engineering: advancing the generation of 'biofuel'. Appl Microbiol Biotechnol 89:1289–1303
- <span id="page-9-8"></span>10. Collins LD, Daugulis AJ (2000) Biodegradation of phenol at high initial concentrations in two-phase partitioning batch and fed-batch bioreactors. Biotechnol Bioeng 55:155–162
- <span id="page-9-5"></span>11. Crawford RL (1981) Lignin biodegradation and transformation. Wiley, New York
- <span id="page-9-6"></span>12. Dence CW, Lin SY (1992) Introduction. In methods in lignin chemistry. Springer Verlag, New York
- <span id="page-9-9"></span>13. Fava F, Armenante PM, Kafkewitz D, Marchetti L (1995) Influence of organic and inorganic growth supplements on the aerobic biodegradation of chlorobenzoic acid. Appl Microbiol Biotechnol 43:171–177
- <span id="page-9-13"></span>14. Holladay JA, Bozell JJ, White JF, Johnson D (2007) Top valueadded chemicals from biomass volume II-results of screening for potential candidates from biorefinery lignin. Pacific Northwest National Laboratory. PNNL-16983
- <span id="page-9-3"></span>15. Juturu V, Wu C (2012) Microbial xylanases: engineering, production and industrial applications. Biotechnol Adv 30:1219–1227
- <span id="page-10-23"></span>16. Kalme S, Jadhav S, Jadhav M, Govindwar S (2009) Textile dye degrading laccase from *Pseudomonas desmolyticum* NCIM 2112. Enzym Microb Technol 44:65–71
- <span id="page-10-0"></span>17. Lauro B, Rossi M, Moracci M (2006) Characterization of β-glycosidase from the thermoacidophilic bacterium *Alicyclobacillus acidocaldarius*. Extremophiles 10:301–310
- <span id="page-10-10"></span>18. Leonard D, Lindley ND (1998) Carbon and energy flux constraints in continuous cultures of *Alcaligenes eutrophus* grown on phenol. Microbiology 144:241–248
- <span id="page-10-6"></span>19. Mahadevaswamy M, Mall ID, Prasad B, Mishra IM (1997) Removal of phenol by adsorption on coal fly ash and activated carbon. Pollut R 16:170–175
- <span id="page-10-25"></span>20. Martins LO, Durao P, Vrissos V, Lindley PF (2015) Laccases of prokaryotic origin: enzymes at the interface of protein science and protein technology. Cell Mol Life Sci 72:911–922
- <span id="page-10-13"></span>21. Mavromatis K, Sikorski J, Lapidus A, Del Rio TG, Copeland A, Tice H, Cheng JF, Lucas S, Chen F, Nolan M, Bruce D, Goodwin L, Pitluck S, Ivanova N, Ovchinnikova G, Pati A, Chen A, Palaniappan K, Land M, Hauser L, Chang YJ, Jeffries CD, Chain P, Meincke L, Sims D, Chertkov O, Han C, Brettin T, Detter JC, Wahrenburg C, Rohde M, Pukall R, Goker M, Bristow J, Eisen JA, Markowitz V, Hugenholtz P, Klenk HP, Kyrpides NC (2010) Complete genome sequence of *Alicyclobacillus acidocaldarius* type strain (104-IAT. Stand Genomic Sci 2:9–18
- <span id="page-10-11"></span>22. Nikakhtari H, Hill GA (2006) Continuous bioremediation of phenol-polluted air in an external loop airlift bioreactor with a packed bed. J Chem Technol Biotechnol 81:1029–1038
- <span id="page-10-24"></span>23. Niladevi K, Sukumaran NRK, Prema P (2007) Utilization of rice straw for laccase production by *Streptomyces psammoticus* in solid-state fermentation. J Ind Microbiol Biotechnol 34:665–674
- <span id="page-10-9"></span>24. Paller G, Hommel RK, Kleber HP (1995) Phenol degradation by *Acinetobacter calcoaceticus* NCIB 8250. J Basic Microbiol 35:325–350
- <span id="page-10-5"></span>25. Perlack RD, Wright LL, Turhollow AF, Graham RL, Stokes BJ, Erbach DC (2005) Biomass as feedstock for a bioenergy and bioproducts industry: the technical feasibility of a billion-ton annual supply. ORNL/TM-2005/66; Oak Ridge National Laboratory: Oak Ridge
- <span id="page-10-7"></span>26. Prpich GP, Daugulis AJ (2005) Enhanced biodegradation of phenol by a microbial consortium in a solid-liquid two-phase partitioning bioreactor. Biodegradation 16:329–339
- <span id="page-10-14"></span>27. Reiss R, Ihssen J, Richter M, Eichhorn E, Schilling B, Thony-Meyer L (2013) Laccase versus laccase-like multi-copper oxidase: a comparative study of similar enzymes with diverse substrate spectra. PLoS One 8:e65633
- <span id="page-10-17"></span>28. Rolfe MD, Rice CJ, Lucchini S, Pin C, Thompson A, Cameron ADS, Alston M, Stringer MF, Betts RP, Baranyi J, Peck MW, Hinton JCD (2012) Lag phase is a distinct growth phase that prepares bacteria for exponential growth and involves transient metal accumulation. J Bacteriol 194:686–701
- <span id="page-10-22"></span>29. Ruijssenaars HJ, Hartmans S (2004) A cloned *Bacillus halodurans* multicopper oxidase exhibiting alkaline laccase activity. Appl Microbiol Biotechnol 65:177–182
- <span id="page-10-8"></span>30. Ruiz-ordaz N, Ruiz-Lagunez JC, Castanou-Gonzalez JH, Hernandez-Manzano E, Cristiani-Urbina E, Galindez-Mayer J (2001) Phenol biodegradation using a repeated batch culture of *Candida tropicalis* in a multistage bubble column. Revista Latinoamericana de Microbiogia 43:19–24
- <span id="page-10-12"></span>31. Shen B, Hutchinson CR (1993) Tetracenomycin F1 monooxygenase: oxidation of a naphthacenone to a naphthacenequinone in the biosynthesis of tetracenomycin C in *Streptomyces glaucescens*. Biochemistry 32:6656–6663
- <span id="page-10-21"></span>32. Solomon BO, Posten C, Harder MPF, Hecht V, Zilli M, Converti A, Lodi A, DelBorghi M, Ferraiolo G, Deckwer WD (1994) Energetics of *Pseudomonas* phenol removal from waste gases with a biological *cepacia* growth in a chemostat with phenol limitation, J. fitter by *Pseudomonas putida*. Biotechnol Bioeng 60:275–282
- <span id="page-10-3"></span>33. ten Have R, Teunissen PJM (2001) Oxidative mechanisms involved in lignin degradation by white-rot fungi. Chem Rev 101:3397–3413
- 34. Thompson DN, Apel WA, Thompson VS, Reed DW, Lacey JA (2011) Thermophlic and thermoacidophilic biopolymer-degrading genes and enzymes from Alicyclobacillus acidocaldarius and related organisms, methods. US Patent 8,507,242, 7th July 2011
- <span id="page-10-4"></span>35. Thompson DN, Apel WA, Thompson VS, Ward TE (2013) Methods of combined bioprocessing and related microorganisms, thermophilic and/or acidophilic enzymes, and nucleic acids encoding said enzymes. US Patent 8,691,525, 14th Nov 2013
- <span id="page-10-19"></span>36. Thompson VS, Thompson DN, Shcaller KD, Apel WA Enzyme and methodology for the treatment of a biomiass. US Patent 7,727,755, 2nd Nov 2005
- <span id="page-10-16"></span>37. Ullrich R, Hofrichter M (2007) Enzymatic hydroxylation of aromatic compounds. Cell Mol Life Sci 64:271–293
- <span id="page-10-20"></span>38. Whiteley CG, Lee DJ (2006) Enzyme technology and biological remediation. Enzyme Microb Tech 38:291–316
- <span id="page-10-18"></span>39. Widsten P, Kandelbauer A (2008) Laccase applications in the forest products industry: a review. Enzyme Microb Technol 42:293–307
- <span id="page-10-15"></span>40. Yamasaki H, Abdel-Ghany SE, Cohu CM, Kobayashi Y, Shikanai T, Pilon M (2007) Regulation of copper homeostasis by micro-RNA in Arabidopsis. J Biol Chem 282:16369–16378
- <span id="page-10-1"></span>41. Zambrano JL, Ettayebi K, Maaty WS, Faunce NR, Bothner B, Hardy ME (2011) Rotavirus infection activates the UPR but modulates its activity. Virol J 8:359–372
- <span id="page-10-2"></span>42. Zhang Y, Ju J, Peng H, Gao F, Zhou C, Zeng Y, Xue Y, Henrissat B, Gao GF, Yanhe M (2008) Biochemical and structural characterization of the intracellular mannanase AaManA of *Alicyclobacillus acidocaldarius* reveals a novel glycoside hydrolase family belonging to Clan GH-A. J Biol Chem 283:31551–31558